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# A New Species Richness Measure Improved from Margalef and Menhinick Indices

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#### **Highlights**

- · A new species richness calculator was proposed as an alternative to Margalef and Menhinick indices.
- Results indicated that the proposed estimator presented better performance.
- New estimator likely yields lower values if species vastly outnumber individuals

#### **Article Info**

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# Keywords

Alpha diversity Assemblage BİÇEB software Species richness Diversity estimators

#### Abstract

In the present study, a new estimator  $\binom{\alpha^*D_{MM}}{\alpha^*D_{MM}}$  is proposed to measure species richness. Its' essential inputs are alfa  $(\alpha)$  scale parameters of Margalef  $(D_{MG})$  and Menhinick  $(D_{MN})$  indices. To evaluate the performance of  $\alpha^*D_{MM}$ , both hypothetical and natural community datasets were used. The computations were performed using a spreadsheet program created for  $\alpha^*D_{MM}$ . The new proposed estimator is also integrated into the BİÇEB software. According to estimation results,  $\alpha^*D_{MM}$  seems to be a better form rather than  $D_{MG}$  and  $D_{MN}$ . Therefore it may be employed for comparing species richness of the natural communities. To better understand the performance of  $\alpha^*D_{MM}$ , further studies should be generated using various types of real ecological data.

# 1. INTRODUCTION

Biodiversity plays a vital role for ecosystem productivity, stability, health and dynamism. It is therefore the main issue in conservation biology, environmental ecology and biogeography [1-2].

As stated by Peet 1974, biodiversity, in essence, has always been defined by using the measures. Various diversity measures have been proposed in the literature. All these measures are broadly divided into three groups which are known as species richness indices, heterogeneity indices and species abundance models [3]. To estimate species diversity using heterogeneity indices and species abundance models, the main inputs are the number of species and their abundance or incidence data. Among those measures, Shannon entropy [4], Simpson diversity index [5] and log-normal [6] are the most popular ones.

In comparison with heterogeneity indices and species abundance models, species richness indices have simpler forms. The core members of species richness indices are species richness (S) [7], Woodwell index  $(R_w)$  [8], Menhinick index  $(D_{MN})$  [9] and Margalef index  $(D_{MG})$  [10], a small modification of Odum index  $(R_o)$  [11]. Those indices estimate species richness without using species abundance or incidence data. Among them, the oldest, simplest and still most commonly used measure is species richness (S). It refers to the number of species present in an area or an assemblage without regard to number of individuals [3]. Unlike S, the other richness indices  $(R_w, D_{MN})$  and  $(R_w, R_w)$  and  $(R_w, R_w)$  assume that there is a relationship between  $(R_w, R_w)$  and  $(R_w, R_w)$  and  $(R_w, R_w)$  have the same structural characteristics,

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the importance degrees to N given by them are different. The weight of N is equal to N in  $R_w$  whereas it corresponds to  $\sqrt{N}$  in  $D_{MN}$  and,  $\ln N$  in  $D_{MG}$  respectively.

 $R_w$  has been scarcely preferred in estimation of species richness. However,  $D_{MN}$  and  $D_{MG}$  have frequently used since they are intuitively much more favorable measures compared to  $R_w$ .

Regarding to  $D_{MN}$  and  $D_{MG}$ , there is no consensus about which one should be used to compare species richness. Therefore, researchers generally use both of those indices instead of selection one of them [12-17]. Nevertheless employment both of them may cause conflicting results in comparisons of species richness among communities. For instance, suppose that 2 communities, A and B. Community A includes 15 species and totally of 45 individuals and, Community B is composed of 20 species and 150 individuals. A is richer than B according to  $D_{MN}$  results ( $D_{MN}(A) = 2.24, D_{MN}(B) = 1.63$ ) whereas  $D_{MG}$  results indicate that B is richer than A ( $D_{MG}(A) = 3.68, D_{MG}(B) = 3.79$ ). As can be understood from this comparative example, selecting both of the indices or one of them is not a solution. The solution may be produced an index from both of them. To the best of our knowledge, there is no such an estimator produced or derived from  $D_{MN}$  or  $D_{MG}$  in the literature.

The present study offers a species richness measure derived from  $D_{MN}$  and  $D_{MG}$  and explains how to employ this measure in estimation of species richness using hypothetical and natural community data.

#### 2. MATERIAL AND METHOD

As mentioned in the introduction, the following equations are the core members of species richness measures

$$S = \sum_{i=1}^{S} i^0 \tag{1}$$

$$D_{MG} = \frac{S-1}{\ln(N)} \tag{2}$$

$$D_{MN} = \frac{S}{\sqrt{N}} \tag{3}$$

$$R_{w} = S/N \tag{4}$$

where S,  $D_{MG}$ ,  $D_{MN}$  and  $R_w$  are species richness [7], Margalef index [10], Menhinick index [9], and Woodwell index [8] respectively. In the equations, N represents total number of individuals.

By using a scale parameter ( $\alpha$ ) Equation (1), Equation (3) and Equation (4) can be combined into one formula ( ${}^{\alpha}DS$ ).

$${}^{\alpha}DS = \frac{S}{N^{\alpha}}, \ 1 \ge \alpha \ge 0 \tag{5}$$

Where the formula reduces  $S\left(\frac{\sum_{i=1}^{S}i^{0}}{N^{0}}\right)$  at  $\alpha=0$ ,  $D_{MN}$  at  $\alpha=0.5$  ( $\alpha_{D_{MN}}=0.5$ ) and  $R_{w}$  at  $\alpha=1$ . Suppose that a community includes a total of 8 species (S=8) and 27 individuals (N=27). Equation (5) allows us to create a curve (Figure 1).

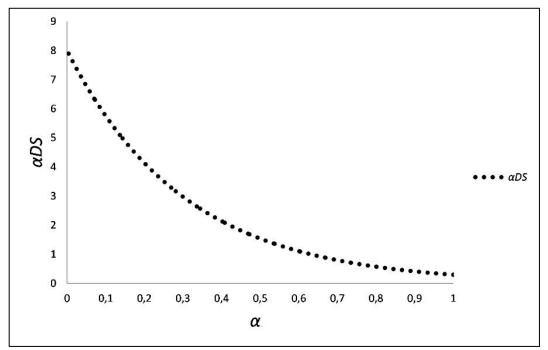


Figure 1. The profile of  $^{\alpha}DS$  for S=8 and N=27 (S=8 ( $\alpha=0$ ) and  $R_{w}=0.296$  ( $\alpha=1$ ))

This curve may be called species richness profile. That is however not the issue focused on. The focal issue is related to  $D_{MG}$ .  $D_{MG}$  has not a fixed  $\alpha$  value because it contains  $\ln(N)$ . Its  $\alpha$  value  $(\alpha_{D_{MG}})$  generally changes from 0 to 0.5 depending on S and N values. Herein the value of  $D_{MG}$  (2.12) corresponds to  $\alpha = 0.4024$  for S = 8 and N = 27 (Figure 2).

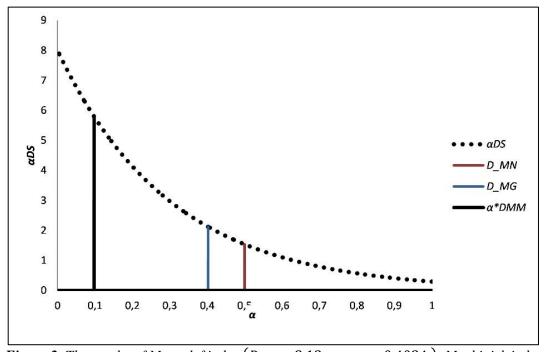


Figure 2. The results of Margalef index ( $D_{MG} = 2.12$ ,  $\alpha_{D_{MG}} = 0.4024$ ), Menhinick index ( $D_{MN} = 1.54$ ,  $\alpha_{D_{MN}} = 0.5$ ) and the proposed estimator ( $\alpha^* D_{MM} = 5.80$ ,  $\alpha^* = 0.0976$ ) for S = 8 and N = 27

The proposed estimator  $(\alpha^* D_{MM})$  is based on  $\alpha$  values of  $D_{MN}$  and  $D_{MG}$ 

$$\alpha^* D_{MM} = c_{2S} \frac{S}{N^{\alpha^*}} \tag{6}$$

where,

$$\alpha^* = \alpha_{D_{MN}} - \alpha_{D_{MG}} c_{1S} = 0.5 - \alpha_{D_{MG}} c_{1S} . \tag{7}$$

 $c_{1S}$  and  $c_{2S}$  are the correction coefficients.  $c_{2S}$  gets the fixed values which are 1, 1.208, 1.258, 1.202, 1.150, 1.080 and 1.016 from S=1 to S=7. The fixed values of  $c_{1S}$  are 0.4, 0.42, 0.54, 0.7, 0.8, 0.9 and 0.98 from S=1 to 7. If S>7, then  $c_{2S}=c_{1S}=1$ . Note that if there is only one species and one individual in a community (S=N=1), it is assumed that  $\alpha^*D_{MM}=1$ . That constraint is due to  $D_{MG}$  numerator (If S=1, S-1=0). For S=8 and N=27, all the computed values are given in Figure 2. A dataset composed of 10 hypothetical communities was created and used to evaluate the performance of the proposed estimator (Table 1).

**Table 1.** S and N values of the hypothetical communities

| Communities | $C_1$ | $C_2$ | $C_3$ | <i>C</i> <sub>4</sub> | <i>C</i> <sub>5</sub> | <i>C</i> <sub>6</sub> | <i>C</i> <sub>7</sub> | <i>C</i> <sub>8</sub> | <i>C</i> <sub>9</sub> | <i>C</i> <sub>10</sub> |
|-------------|-------|-------|-------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|-----------------------|------------------------|
| S           | 15    | 15    | 15    | 15                    | 20                    | 20                    | 20                    | 20                    | 15                    | 25                     |
| N           | 15    | 30    | 45    | 930                   | 20                    | 150                   | 1800                  | 70000                 | 20                    | 150                    |

The natural community data obtained from the Sütçüler District of Isparta province in the Mediterranean Region were used to evaluate the new proposed estimator [18]. In the mentioned study, plant cover data were recorded using Braun-Blanquet cover-abundance scale. Then, the plant cover data were transformed according to Westhoff and Maarel [19] so that each corresponding to a value between 1 and 9. Estimations were performed for the data obtained from 10 sample areas. *S*, *N* and the transformed values are given in Table 2. As can be seen in Table 2, a total of 26 taxa were observed in 10 sample plots.

**Table 2.** Transformed abundance values of natural community data

|  | $rC_1$ | $rC_2$ | $rC_3$ | $rC_4$ | $rC_5$ | $rC_6$ | $rC_7$ | $rC_8$ | $rC_9$ | $rC_{10}$ |
|--|--------|--------|--------|--------|--------|--------|--------|--------|--------|-----------|
| Berberis crataegina DC.                | 3      | 3      | 0      | 0      | 0      | 0      | 0      | 2      | 1      | 2         |
| Cistus salviifolius L.                 | 2      | 2      | 5      | 3      | 2      | 3      | 1      | 3      | 3      | 5         |
| Colutea cilicica Boiss. & Balansa      | 0      | 0      | 2      | 2      | 0      | 0      | 0      | 0      | 0      | 0         |
| Crataegus orientalis Pallas ex Bieb.   | 3      | 2      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Daphne gnidioides Jaub. & Spach        | 0      | 0      | 0      | 2      | 0      | 0      | 0      | 0      | 0      | 0         |
| Daphne sericea Vahl.                   | 0      | 0      | 3      | 3      | 2      | 2      | 2      | 3      | 3      | 0         |
| Fontanesia phillyreoides Labill.       | 0      | 0      | 1      | 2      | 3      | 3      | 2      | 2      | 0      | 0         |
| Jasminium fructicans L.                | 0      | 0      | 1      | 2      | 2      | 2      | 0      | 0      | 2      | 2         |
| Juniperus excelsa Bieb.                | 0      | 0      | 0      | 0      | 0      | 0      | 5      | 5      | 7      | 5         |
| Juniperus oxycedrus L.                 | 2      | 1      | 5      | 5      | 5      | 3      | 0      | 0      | 0      | 0         |
| Origanum minutiflorum Schwrd et Davis. | 0      | 2      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Origanum onites L.                     | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Osyris alba L.                         | 0      | 0      | 0      | 0      | 0      | 0      | 1      | 0      | 0      | 0         |
| Paliurus spina-christi Mill.           | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Phillyrea latifolia L.                 | 0      | 0      | 5      | 5      | 5      | 5      | 5      | 5      | 5      | 5         |
| Phlomis grandiflora H.S. Thompson.     | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 2         |
| Pinus brutia var. brutia Ten.          | 0      | 0      | 7      | 7      | 5      | 7      | 8      | 7      | 5      | 5         |
| Pinus nigra J.F.Arnold                 | 8      | 8      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Pistacia terebinthus L.                | 0      | 0      | 3      | 2      | 3      | 2      | 5      | 5      | 5      | 5         |
| Quercus cercis L.                      | 2      | 2      | 0      | 0      | 1      | 2      | 0      | 0      | 0      | 0         |
| Quercus coccifera L.                   | 0      | 0      | 8      | 5      | 5      | 5      | 7      | 7      | 3      | 5         |
| Rosa canina L.                         | 2      | 3      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Ruscus aculeatus L.                    | 0      | 0      | 0      | 0      | 1      | 2      | 0      | 0      | 0      | 0         |
| Styrax officinalis L.                  |        | 0      | 1      | 0      | 0      | 3      | 3      | 3      | 0      | 2         |
| Teucrium chamaedrys L.                 | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0      | 0         |
| Teucrium polium L.                     | 0      | 0      | 2      | 1      | 0      | 2      | 0      | 0      | 0      | 0         |

### 3. RESULTS AND DISCUSSION

As can be seen in Table 1, the first four communities  $(C_1: C_4)$  have the same S value (S=15) but different N values changing from 15 to 930. From  $C_1$  to  $C_4$ , with decreasing  $D_{MN}$  and  $D_{MG}$ ,  $\alpha^*D_{MM}$  decreases.  $D_{MN}$ ,  $D_{MG}$  and  $\alpha^*D_{MM}$  have similar trends from  $C_5$  to  $C_8$  since they have the same number of species (S=20) with different N values ranging from 20 to 70000. Parallel trends of  $D_{MN}$ ,  $D_{MG}$  and  $\alpha^*D_{MM}$  are in line with expectation for a fixed S value agains changing N values. As a result of comparisons within group 1  $(C_1: C_4)$  and within group 2  $(C_5: C_8)$ , the difference of  $\alpha^*D_{MM}$  from  $D_{MN}$  and  $D_{MG}$  is invisible.

The difference of  $\alpha^*D_{MM}$  can however be understood by comparing  $C_9$  with  $C_{10}$  and  $C_4$  with  $C_7$ .  $C_{10}$  contains more species and much more individuals than  $C_9$  (Table 3). According to  $D_{MN}$  results,  $C_9$  is richer than  $C_{10}$ . However,  $D_{MG}$  results indicate that  $C_{10}$  is a richer community than  $C_9$ . As expected, the decision of  $\alpha^*D_{MM}$  is in favor of  $D_{MN}$  because the occurrence probability of the number of rare species in particular singletons is likely to be much bigger in  $C_9$  rather than  $C_{10}$ .

**Table 3.** The species richness results of hypothetical communities ( $\alpha_{D_{MN}} = 0.5$ )

| Communities | $\alpha_{D_{MG}}$ | $\alpha^*$ | $D_{MN}$ | $D_{MG}$ | $^{lpha^*}D_{MM}$ |
|-------------|-------------------|------------|----------|----------|-------------------|
| $C_1$       | 0.3934            | 0.1066     | 3.8730   | 5.1698   | 11.2388           |
| $C_2$       | 0.3802            | 0.1198     | 2.7386   | 4.1162   | 9.98003           |
| $C_3$       | 0.3693            | 0.1307     | 2.2361   | 3.6778   | 9.12045           |
| $C_4$       | 0.2913            | 0.2087     | 0.4919   | 2.0482   | 3.60221           |
| $C_5$       | 0.3834            | 0.1166     | 4.4721   | 6.3424   | 14.1036           |
| $C_6$       | 0.3319            | 0.1681     | 1.6330   | 3.7919   | 8.61448           |
| $C_7$       | 0.2756            | 0.2244     | 0.4714   | 2.5348   | 3.72003           |
| $C_8$       | 0.2208            | 0.2792     | 0.0756   | 1.7031   | 0.88773           |
| $C_9$       | 0.3893            | 0.1107     | 3.3541   | 4.6733   | 10.7663           |
| $C_{10}$    | 0.3298            | 0.1702     | 2.0412   | 4.7898   | 10.6554           |

When comparing  $C_4$  to  $C_7$ , we see that there is no agreement between  $D_{MN}$  and  $D_{MG}$  once again.  $\alpha^*D_{MM}$  supports  $D_{MG}$  results this time. The decision of  $\alpha^*D_{MM}$  is intuitively accurate since both  $C_4$  and  $C_7$  have a great number of total individuals comparing to their number of species. This means that none of them is most likely to include a notable negative bias. Consequently,  $\alpha^*D_{MM}$  seems to reduce to the effect of negative bias originated from occurrences of rare species in a community. However, it does not mean that the results of  $\alpha^*D_{MM}$  correspond to true species richness because true species richness is equal or larger than the observed number of species.

S and N values estimated for the natural community data of 10 sample plots obtained from Sütçüler District are given in Table 3. To avoid confusion with the hypothetical data, the sample plots in the natural community data were named  $rC_1...rC_{10}$  from 1 to 10.

Table 4. S and N values of natural community data

| Communities | $rC_1$ | $rC_2$ | $rC_3$ | $rC_4$ | $rC_5$ | $rC_6$ | $rC_7$ | $rC_8$ | $rC_9$ | $rC_{10}$ |
|-------------|--------|--------|--------|--------|--------|--------|--------|--------|--------|-----------|
| S           | 7      | 8      | 12     | 12     | 11     | 13     | 10     | 10     | 9      | 10        |
| N           | 22     | 23     | 43     | 39     | 34     | 41     | 39     | 42     | 34     | 38        |

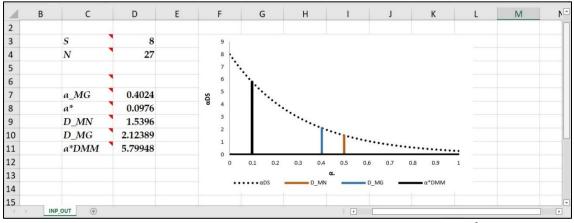
After obtaining S and N values,  $\alpha_{D_{MG}}$ ,  $\alpha^*$ ,  $D_{MN}$ ,  $D_{MG}$  and  $\alpha^*D_{MM}$  estimations were performed (Table 5).

| Communities | $\alpha_{D_{MG}}$ | $\boldsymbol{lpha}^*$ | $D_{MN}$ | $D_{MG}$ | $^{\alpha^*}D_{MM}$ |
|-------------|-------------------|-----------------------|----------|----------|---------------------|
| $rC_1$      | 0.415             | 1.4924                | 0.0933   | 1.941    | 5.3327              |
| $rC_2$      | 0.4071            | 1.6681                | 0.0929   | 2.2325   | 5.9784              |
| $rC_3$      | 0.3743            | 1.83                  | 0.1247   | 2.9246   | 7.5074              |
| $rC_4$      | 0.3782            | 1.9215                | 0.1218   | 3.0025   | 7.6805              |
| $rC_5$      | 0.3844            | 1.8865                | 0.1156   | 2.8358   | 7.3173              |
| $rC_6$      | 0.3749            | 2.0303                | 0.1251   | 3.2314   | 8.1693              |
| $rC_7$      | 0.3832            | 1.6013                | 0.1168   | 2.4566   | 6.5187              |
| $rC_8$      | 0.3809            | 1.543                 | 0.1191   | 2.4008   | 6.4072              |
| $rC_9$      | 0.3908            | 1.5435                | 0.1092   | 2.2686   | 6.1236              |
| $rC_{10}$   | 0.384             | 1.6222                | 0.116    | 2.4762   | 6.5576              |

**Table 5.** The species richness results of natural community data ( $\alpha_{DMN} = 0.5$ )

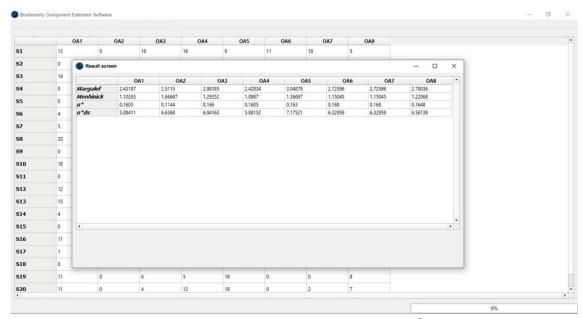
When the values in Tables 4 and 5 are examined, it is seen that the S values of  $rC_3$  and  $rC_4$  are the same. However, the N value of  $rC_3$  (43) is higher than the N value of  $rC_4$  (39). Nevertheless, in terms of  $D_{MN}$ , the value of  $rC_3$  is higher than that of  $rC_4$ , while for  $D_{MG}$  the value of  $rC_4$  is higher.  $\alpha^*D_{MM}$  is higher in  $rC_4$ , which has a lower N value. Similar results are also observed for  $rC_7$  and  $rC_8$ . Özkan [3] obtained different results for  $D_{MN}$  and  $D_{MG}$  in his calculations based on hypothetical data. However, it is thought that these results alone cannot be an indicator. Since, Süel et al. [20] estimated higher values for DMG in all 43 sample areas in their study. Mulya et al. [21] also compared species richness and diversity estimations in their study. Researchers stated that  $D_{MG}$  produced different results and performed better than Shannon, Simpson,  $D_{MN}$  in terms of their data. In contrast to this study, Davari et al. [13] found that  $D_{MN}$  performed better than  $D_{MG}$ . Both the aforementioned studies and our study shown that non-overlapping results can be observed in different data in terms of  $D_{MN}$  and  $D_{MG}$ . Therefore,  $\alpha^*D_{MM}$  is suggested as a useful option.

Lastly, it is worth mentioning here that  $\alpha^*D_{MM}$  may take a smaller value than  $D_{MG}$  as can be seen the values of  $C_8$  in Table 2. The lower limit  $\alpha^*DS$  goes to the value of  $D_{MN}$  and, S draws the line of its' upper limit. The spreadsheet program ( $DMM_macro_v1$ ) that calculates the parameters in Table 2 is available on the website at <a href="https://kantitatifekoloji.net/takdivozkan">https://kantitatifekoloji.net/takdivozkan</a>. Application of the spreadsheet program is simple. Enter the values of S and N into D3 and D4 cells, respectively. The outputs ( $\alpha_{D_{MG}}$ ,  $\alpha^*$ ,  $D_{MN}$ ,  $D_{MG}$  and  $\alpha^*D_{MM}$ ) appear in D7-D11 cells (Figure 3).



**Figure 3.** The menu of the spreadsheet program for computation of  $\alpha^*D_{MM}$ 

In addition, related calculations are integrated into the BİÇEB software [22] for the calculation of the parameter. BİÇEB is based on python and a version suitable for different operating systems (Mac, Windows, and Linux) that can be downloaded free of charge from the https://kantitatifekoloji.net/biceb (Figure 4). BİÇEB is a software developed using Python programming language within the scope of TÜBİTAK 1005 grant. It can be downloaded free of charge from the link provided. Since it is developed in open code, it can be used free of charge on all operating systems (Windows, Mac OS and Linux). When the program is downloaded, it has the potential to be easily used by researchers since it comes with a download and user manual.



**Figure 4.** BİÇEB software for computation of  $\alpha^*D_{MM}$ 

#### 4. CONCLUSIONS

The present study offers a new richness estimator  $(\alpha^*D_{MM})$ .  $\alpha_{D_{MG}}$  and  $\alpha_{D_{MN}}$  are the scale values forming the basis of  $\alpha^*D_{MM}$  (Equation (6)). With increasing proportional differences between S and N,  $\alpha^*D_{MM}$  value decreases. If there is an enormous difference between S and S, it is very likely that  $\alpha^*D_{MM}$  takes a less value than  $D_{MG}$ . The essential difference of  $\alpha^*D_{MM}$  arises by comparing community pairs such as  $C_9$  with  $C_{10}$  or  $C_4$  with  $C_7$  given in Table 1.

The results obtained from the data of the hypothetical and natural communities indicate that the new proposed index have the ability to represent both of D\_MG and D\_MN. That is why it seems to promising to estimate species richness. However, to better understand strengths and weaknesses of the proposed estimator, further studies should be performed using natural community data. In addition, it is thought that the new estimator can be preferred by researchers due to both the development of MS Excel Macro and the possibility of free use through the BİÇEB software.

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# **CONFLICTS OF INTEREST**

No conflict of interest was declared by the authors.

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